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Genotypic characteristics of multidrug-resistant *Pseudomonas aeruginosa* from hospital wastewater treatment plant in Rio de Janeiro, Brazil

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Keywords

clonal complexes, hospital sewage, multidrugresistant, multilocus sequence typing, *Pseudomonas aeruginosa*.

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Introduction

Abstract

Aims: To investigate *Pseudomonas aeruginosa* isolates from a hospital wastewater treatment plant (HWTP), focusing on enzyme-based mechanisms of β -lactams resistance and the genetic relatedness among isolates.

Methods and Results: Forty-one *Ps. aeruginosa* strains recovered from a HWTP were identified by amplification of 16S rRNA gene. β -lactamase production was screened by disc diffusion, CHROMagar extended-spectrum β -lactamase (ESBL) and β -lactamase strips. β -lactamase and ESBL producing isolates were investigated by PCR for the presence of ESBL, metallo- β -lactamase and *Klebsiella pneumoniae* carbapenemase encoding genes. Thirty-four isolates (83%) were resistant to at least one antibiotic belonging to three or more classes. Out of these 34 isolates, 28 (82%) were classified as multidrug-resistant (MDR) and 6 (18%) extensively drug-resistant (XDR). Genetic relatedness by Enterobacterial Repetitive Intergenic Consensus sequence-PCR and Multilocus sequence typing analysis showed 20 distinct profiles and 15 sequencing types respectively. Clonal Complex 244 (CC244) shows the pathogenic potential of this clone carrying MDR and XDR strains from clinical, environmental and hospital waste sources.

Conclusions: Our results suggest that treatment facilities for hospital wastewater can stimulate the increase of antimicrobial resistance bacteria and genes.

Significance and Impact of the Study: The great genetic diversity of *Ps. aeruginosa* recovered from HWTP constantly released into aquatic systems allow the spread of antimicrobial-resistant organisms and genes.

Pseudomonas aeruginosa is an opportunistic pathogen with intrinsic resistance to many antimicrobials. Furthermore, under selective pressure, this micro-organism may easily develop powerful resistance either by mutation in chromosomally encoded genes or by horizontal transfer of resistance genes (Zhao and Hu 2010). Infections caused by this pathogen are often difficult to treat because of its multidrug-resistant (MDR) phenotype (De Francesco *et al.* 2013). Production of β -lactamases, such as extended-spectrum β -lactamases (ESBLs) and carbapenemases, is an important mechanism of β -lactam antibiotics resistance in *Ps. aeruginosa* nosocomial isolates, which jeopardizes antimicrobial therapy in hospitalized patients (Picão *et al.* 2009). Carbapenemase production is of particular concern as it confers resistance to all β -lactams including extended-spectrum cephalosporins, monobactams and carbapenems (Zavascki *et al.* 2010). Carbapenemases that were identified in *Ps. aeruginosa* so far include: metallo- β -lactamases (MBLs) of the VIM (Verona imipenemase), IMP (imipenemase), SPM (São Paulo metallo β -lactamase), GIM (German imipenemase) AIM (Australian imipenemase), DIM (Dutch imipenemase) and NDM types (New Delhi metallo- β -lactamase); KPC (*Klebsiella pneumoniae* carbapenemase), GES (Guiana extended spectrum) types (GES-2, -4, -5, -6 and -11) (Poirel *et al.* 2002; Gupta 2008; Jovcic *et al.* 2011) and OXA (Oxacillinase) type carbapenemases (Poirel *et al.* 2007, 2010). Among these resistance mechanisms, SPM-1, VIM-2, IMP-1, IMP-16, GES-5 and KPC-2 (Toleman *et al.* 2002; Mendes *et al.* 2004; Sader *et al.* 2005; Martins *et al.* 2007; Picão *et al.* 2009; Silva *et al.* 2011; Jácome *et al.* 2012) were found in *Ps. aeruginosa* from Brazil, most of which were clinical isolates.

Hospital effluents may represent a great risk to human public health due to the presence of pathogens and chemicals such as disinfectants, anaesthetics, heavy metals, antimicrobial agents and other drugs that are not metabolized by patients (Emmanuel *et al.* 2005). The disposal of this liquid waste in the treatment of hospital wastewater treatment plant (HWTP) can favour, the selective pressure, an increase in bacterial populations with phenotypes of multidrug resistance to antimicrobials (Prado *et al.* 2008).

Many carbapenem-resistant *Ps. aeruginosa* isolates and clinically important ESBLs-and carbapenemase-encoding genes have been found in hospital wastewater (Santoro *et al.* 2012), hospital and municipal sewage (Korzeniewska *et al.* 2013), urban sewerage systems (containing rainwater, hospital and urban wastewater) (Slekovec *et al.* 2012) wastewater contaminated rivers (Fontes *et al.* 2011; Amos *et al.* 2014). Of notice, NDM-1-producing *Ps. aeruginosa* was identified in tap water collected in India (Walsh *et al.* 2011).

In order to circumvent this problem, hospital effluents have been subjected to different treatments. However, several studies have showed that antibiotic resistance bacteria and genes persist in effluents of a variety of full-scale managements at levels above those typical of aquatic environments, even after disinfection (Auerbach *et al.* 2007).

The aim of this study was to investigate the antimicrobial susceptibility profiles and genetic relatedness of *Ps. aeruginosa* isolates obtained from a HWTP. We also explored the β -lactam-encoding genes and the clonal relationships among the isolates using Multilocus sequence typing (MLST) scheme.

Materials and methods

Study setting and sewage sampling

The effluent treatment system studied is located at a healthcare complex in the metropolitan area of Rio de

Janeiro city, Brazil. The facility holds 322 beds and receives around 30 000 persons per month. The hospital contains a wastewater treatment plant that treats 220 cubic meters of sewage per day. It performs tertiary treatment in four stages, including pretreatment to remove gross solids; an aeration tank (continuous stirred tank reactor with sludge recycle), a clarifier tank and a posttreatment (disinfection of final effluent by chlorination). The treated effluent is discharged in rainwater network and then, into water bodies such as rivers and seawater.

Samples were collected in 2008 and in 2010 from five steps across the station (Grid-affluent, aeration tank, sludge, chlorination tank and treated effluent). Five hundred millilitres per site were collected approximately 20 cm below the surface in sterile glass bottles. Samples were kept refrigerated (4°C) until processed in the laboratory, within 8 h of collection. Temperature, pH, conductivity, dissolved oxygen, turbidity and salinity was assessed at the time of sample collection using Water Quality Checker U-10 (HORIBA, Kyoto, Japan) and dosage of chlorine was carried out using chlorine meter (Homis, São Paulo, Brazil).

Isolation and identification of Pseudomonas aeruginosa

A 100 ml-aliquot of each sample was filtered on membranes of 0.45 μ m porosity. The membranes were transferred to selective asparagine broth and were incubated at 37°C for 24 h. The appearance of green fluorescence under ultraviolet light was considered positive for Ps. aeruginosa growth. Aliquots of 1 ml from these tubes were transferred to acetamide broth and were incubated at 37°C for 24 h. Tubes in which the development of pink colour was observed had an aliquot streaked on acetamide broth (Fuentefria et al. 2011). After incubation at 37°C for 24 h, six colonies per plate were selected for further studies, comprising a total of 60 isolates. Isolates were identified phenotypically using Vitek 2 GNI cards following the manufacturer's recommendations (bioMerieux, Marcy L'etoile, France) and results were certified by Ps. aeruginosa-specific PCR analysis (Spilker et al. 2004).

Antimicrobial susceptibility and MDR/XDR classification

The isolates were tested for antibiotic susceptibility by disc diffusion according to the Clinical Laboratory Standard Institute (CLSI 2014). *Escherichia coli* (ATCC 25922) and *Ps. aeruginosa* (ATCC 27853) were used as quality control strains. Seventeen antimicrobial agents were tested: amikacin (30 μ g), aztreonam (30 μ g), cefepime (30 μ g), cefotaxime (30 μ g), ceftazidime (30 μ g), ceftriaxone (30 μ g), ciprofloxacin (5 μ g) gentamicin (10 μ g), imipenem (10 μ g), meropenem (10 μ g), levofloxacin (5 μ g), piperacillin/tazobactam (100/10 μ g), ticarcillin/clavulanic acid (75/10 μ g), tobramycin (10 μ g), polymyxin B (300 units), colistin (10 μ g) and fosfomycin (200 μ g).

The isolates were classified as MDR and extensively drug-resistant (XDR) according to Magiorakos *et al.* (2012). Production of β -lactamases and ESBLs was investigated by β -lactamase strip (Probac, São Paulo, Brazil) and CHROMagar—ESBL (CHROMagar, Paris, France) respectively. Statistical significance (P < 0.05) between resistance profiles of each step were inferred by Fisher or chi-square test.

Detection of ESBL, MBL and KPC genes

The β -lactamase and ESBL positive isolates were screened by PCR for the presence of ESBL (bla_{PER} , bla_{VEB} , bla_{SHV} , $bla_{CTX-M-1}$, $bla_{CTX-M-2}$, $bla_{CTX-M-8}$, $bla_{CTX-M-9}$, $bla_{CTX-M-25}$, bla_{TEM} and bla_{GES}), MBL (bla_{IMP} , bla_{VIM} , bla_{SPM} and bla_{NDM}) and KPC (bla_{KPC}) encoding genes, as previously described (Toleman *et al.* 2002; Nagano *et al.* 2004; Dubois *et al.* 2005; Poirel *et al.* 2011; Doyle *et al.* 2012; Monteiro *et al.* 2012; Wang *et al.* 2012; Ahmed *et al.* 2013; Barguigua *et al.* 2013).

ERIC typing analysis

Genotypic analysis of the strains was investigated by amplification of the Enterobacterial Repetitive Intergenic Consensus sequence (ERIC-PCR). The extraction of genomic DNA was performed using the protocol for Gram-negative Dnaeasy®Blood&Tissuet (Qiagen[®], Valencia, CA, USA) according to the manufacturer's instructions and the primer ERIC2 were used for the amplification as previously described (Versalovic et al. 1991). The amplicons were analysed by gel electrophoresis and stained with ethidium bromide (3 mg ml^{-1}) . The gel was photographed and analysed using ImageQuant300 (GE, Oppsala, Sweden). Band patterns were analysed with BIONUMERICS ver. 6.6 (Applied Maths, Kortrijk, Belgium) using the Dice coefficient and unweighted pair group method with arithmetic mean pair group method with arithmetic average. Isolates with 100% level of similarity were considered clonally related.

Multilocus sequence typing

The ancestral relationship of the *Ps. aeruginosa* isolates were analysed by MLST with a representative strain of each ERIC type. Most primers used for amplification and sequencing of seven housekeeping genes (*acsA*, *aroE*, *guaA*, *mutL*, *nuoD*, *ppsA* and *trpE*) were designed (L. Cacci, personal communication) with primer-blast assistance (http://www.ncbi.nlm.nih.gov/tools/primer-blast) with exception of *acsA*-F (Curran *et al.* 2004). The ampli-

fication of the housekeeping genes was performed as previously described and the nucleotide sequences were sequenced with fluorescent terminators (BigDye; Applied Biosystems, Foster City, CA) on an Applied Biosystems ABI Prism 3730 automated DNA sequences in duplicate for each primer (Otto *et al.* 2008).

Nucleotide sequences were submitted to the MLST database to determine the allelic numbers and Sequence Types (ST). Association of related ST to form clonal complexes (CCs) based on the number of identical alleles shared by ST is not a well-defined rule and can be subjective. The *Ps. aeruginosa* MLST database describes only two CCs: PA01 associated with ST-549 and PA14 associated with ST-253 and no information is disclosed of how these CCs were formed as there is only one ST for each CC. Maâtallah *et al.* (2013) considered that ST sharing five or more identical alleles are part of the same CC.

The MLST profiles were clustered with the BIONUMERICS 6.6 software (Applied Maths, Sint-Martens-Latem, Belgium) using a categorical coefficient and graphing was assessed using the minimum spanning tree tool, as described before (Schouls *et al.* 2004).

Results

Physical and chemical parameters

The pH of the five steps of the first sample analysed was maintained between 6.0-6.8, while in the second sample between 7.6 and 8.4, thus within the range of pH that enables the growth of many micro-organisms, including Ps. aeruginosa. Turbidity showed high values in step 1 and step 2 both samplings, mainly due to the presence of suspended solids in the water. The dissolved oxygen levels had the highest in step 2 where aeration occurs and elevates the levels of Dissolved Oxygen (DO), which is essential for organic matter biodegradation. The temperature of the five steps of two samplings remained at 28-30°C favouring mesophilic forms present. The chlorine concentration was high enough at the step 4 which the liquid chlorine is added to the tank. Mainly due to the presence of hypochlorite ion in high concentration, the conductivity was also high at this point that is directly proportional to the ionization of substances dissolved in the liquid (Table 1).

Pseudomonas aeruginosa identification and antimicrobial susceptibility profile

The five steps of sewage treatment at the HWTP yielded 60 isolates, 41 of which were identified as *Ps. aeruginosa*: step 1 – Grid/affluent (n = 6); step 2 – aeration tank (n = 9); step 3 – sludge (n = 10); step 4 – chlorination tank (n = 4); step 5 – treated effluent (n = 12).

Disc-diffusion susceptibility testing indicated high resistance to various antimicrobial agents. Highest resistance rate was observed for fosfomycin (88%) followed by ticarcillin/clavulanic acid (71%), ceftriaxone (63%), aztreonam (59%), cefotaxime (54%), imipenem (39%), cefepime (24%), meropenem (22%), ceftazidime (20%), tobramycin (20%), polymyxin B (20%), colistin (17%), gentamicin (15%), ciprofloxacin (12%), levofloxacin (12%), piperacillin/tazobactam (12%) and amikacin 2%). Susceptibility levels to fosfomycin, tobramycin, levofloxacin, ciprofloxacin and piperacillin/tazobactam were lower for the first three steps of effluent treatment (grid, aeration and sludge), with significant increase (P < 0.05) on the fourth and fifth steps (chlorination and treated effluent).

Thirty-four out of 41 *Ps. aeruginosa* (83%) presented multidrug resistance profiles and were classified as MDR (82%) and XDR (18%) distributed along the five-step points of the HWTP. Thirty-eight isolates (93%) produced β -lactamase enzymes with 29 (76.3%) ESBL (Fig. 1).

ESBL, KPC and MBL-encoding genes

Among 29 (71%) isolates exhibiting ESBL phenotype, 13 harboured bla_{TEM} (45%), 7 harboured bla_{SHV} (24%) and 1 harboured $bla_{\text{CTX-M-1}}$ (3.5%). It is important to emphasize, however, that we did not sequence the amplicons obtained, thus temoniera (TEM) and sulfhydryl variable (SHV) production may not necessarily confer the ESBL phenotype observed. Thirty-eight (93%) isolates showed antimicrobial resistance for carbapenems, 14 carried bla_{KPC} (37%), 14 (37%) presented bla_{VIM} and 6 presented bla_{SPM} (18%) (Table 2).

ERIC and MLST genotyping

ERIC-PCR was able to fingerprint and assign a specific profile to all *Ps. aeruginosa* strains studied from different steps from a HWTP. Among all studied strains, ERIC-PCR

Table 1 Physical and chemical parameters of HWTP steps

resolved 20 genotypes. The banding patterns were highly reproducible under visual and automated analysis (Fig. 2).

MLST analysis revealed a total of 15 different ST, with five previously described (ST-238, ST-244, ST-381, ST-595 and ST-1621) and 10 new ST which were deposited in the Ps. aeruginosa MLST database (ST-1853, ST-1854, ST-1855, ST-1856, ST-1857, ST-1858, ST-1859, ST-1860, ST-1861, ST-1862) (Table 3). Discriminatory power of ERIC-PCR was higher than MLST as strains belonging to the same ST showed different ERIC profiles. Non-MDR (resistant less than three antibiotics) isolates belonged to single ST (238 and 381), with the exception of ST-244, ST-1621 and ST-1862. ST-244 was present in four of the five steps of HWTP, while the new ST-1859 was present in two steps. These two ST were represented by the larger number of isolates for each (n = 8) (Fig. 3). The 41 isolates were grouped into 15 different ST, with 10 newly described. Among the new ST, nine were associated with at least one MDR strain and one strain belonging to ST-1854 showed a XDR pattern. Other XDR strains were classified as ST-244 and ST-595 and were deposited in the Ps. aeruginosa MLST database.

Discussion

We investigated *Ps. aeruginosa* isolates from the HWPT and determined their antibiotic resistance profile, focusing in particular on enzyme-based mechanisms of resistance to β -lactams and in the determination of the genetic relatedness among isolates. Our results may suggest that wastewater treatment is followed by an increase in resistance profiles, although a quantitative study should be conducted to confirm this hypothesis. Nevertheless, it should be emphasized that Chagas *et al.* (2011) reported 48 and 21% of resistance to cefotaxime and ciprofloxacin, respectively, in Enterobacteriaceae isolates from hospital wastewater. Similar findings regarding *E. coli* from hospital wastewater being resistant to quinolones and cephalosporins have been

| Site (step) | Abiotic | Abiotic parameters | | | | | | | | | | | | | |
|----------------------|---------|--------------------|---------------------------------------|------|------------------|------|--------------------------|------|-------------------|------|---------------|------|-----------------|------|--|
| | pН | | Conductivity μ S cm ⁻¹ | | Turbidity NTU | | DO mg l ⁻¹ | | Temperature °C | | Salinity % | | Chlorine ppm | | |
| | 2008 | 2010 | 2008 | 2010 | 2008 | 2010 | 2008 | 2010 | 2008 | 2010 | 2008 | 2010 | 2008 | 2010 | |
| Grid (1) | 6.8 | 7.6 | 0.39 | 0.82 | 18 | 10 | 4.4 | 4.5 | 29 | 28 | 0.0 | 0.0 | 1 | 0.0 | |
| Aeration (2) | 6.1 | 7.9 | 0.31 | 0.38 | 26 | 99 | 7.6 | 9.2 | 30 | 29 | 0.0 | 0.0 | 1 | -1 | |
| Sludge (3) | 6.1 | 8.2 | 0.29 | 0.38 | 14 | 6 | 1.5 | 9.3 | 30 | 29 | 0.0 | 0.0 | 1 | -1 | |
| Chlorination (4) | 6.3 | 8.4 | 0.52 | 0.48 | 12 | 4 | 3.5 | 9.2 | 30 | 29 | 0.0 | 0.0 | 10 | 10.9 | |
| Treated Effluent (5) | 6.0 | 8.4 | 0.29 | 0.35 | 13 | 7 | 3.4 | 3.7 | 30 | 29 | 0.0 | 0.0 | 1 | 0.01 | |

 μ S cm⁻¹, micro-Siemens per centimetre; NTU, Nephelometric Turbidity Unit; mg l⁻¹ Milligrams per litre; °C, Celsius degrees; %, Per cent; ppm, Parts per million; HWTP, hospital wastewater treatment plant.



Figure 1 Susceptibility of isolates of the five steps of hospital wastewater treatment plant. Filled square, antimicrobial resistance; empty square, antimicrobial susceptibility; β , Beta-lactamase producers; *, extended-spectrum β -lactamase producers; M, multidrug-resistance; X, extensively drug-resistant.

reported among others from Denmark (Jakobsen *et al.* 2008), China (Han *et al.* 2012) and Polland (Korzeniewska and Harnisz 2013). Thus, HWTPs could be a hot spot for antibiotic-resistant bacteria selection (Pruden *et al.* 2013).

Table 2 β -lactamase genes detected in all the isolates of *Pseudomonas aeruginosa*

| Isolates | β -lactamase genes | Isolation step |
|----------|---|----------------|
| P3229 | bla_{SHV} , bla_{KPC} | 1 |
| P3233 | _ | 2 |
| P3234 | bla _{TEM} | 3 |
| P3238 | bla _{TEM} , bla _{SPM} , bla _{VIM} | 4 |
| P3239 | bla _{TEM} , bla _{VIM} | 4 |
| P3240 | bla _{TEM} , bla _{KPC} , bla _{VIM} | 5 |
| P3241 | _ | 5 |
| P3242 | _ | 5 |
| P3246 | bla _{тем} | 5 |
| P3249 | bla _{shv} , bla _{vim} | 2 |
| P3252 | bla_{SHV} , bla_{KPC} , bla_{VIM} | 4 |
| P3671 | _ | 1 |
| P3672 | Ыа _{spm} | 1 |
| P3673 | bla _{SPM} | 1 |
| P3674 | bla _{shv.} bla _{spm} | 1 |
| P3675 | bla _{VIM} | 1 |
| P3676 | bla _{TEM} | 2 |
| P3677 | $bla_{\rm KPC}$, $bla_{\rm VIM}$ | 2 |
| P3678 | bla _{SHV} , bla _{CTX-M-1} , bla _{KPC} | 2 |
| P3679 | bla _{кPC} | 2 |
| P3680 | bla _{кPC} | 2 |
| P3681 | bla _{кPC} | 2 |
| P3682 | Ыа _{крс} | 3 |
| P3683 | _ | 3 |
| P3684 | _ | 3 |
| P3685 | _ | 3 |
| P3686 | bla _{TEM} | 3 |
| P3687 | bla _{кPC} | 3 |
| P3688 | bla _{кPC} | 3 |
| P3689 | bla _{тем} , bla _{кPC} | 3 |
| P3691 | bla _{тем} | 5 |
| P3692 | bla _{TEM} , bla _{VIM} | 5 |
| P3693 | bla _{VIM} | 5 |
| P3694 | bla _{VIM,} bla _{SPM} | 5 |
| P3695 | bla _{VIM,} bla _{SPM} | 5 |
| P3696 | _ | 5 |
| P3697 | _ | 5 |
| P3846 | bla _{TEM} , bla _{SHV} , bla _{KPC} , bla _{VIM} | 2 |
| P3847 | bla _{кPC,} bla _{VIM} | 3 |
| P3848 | bla _{TEM} , bla _{VIM} | 4 |
| P3849 | bla _{TEM} | 5 |

Isolation steps: (1) Grid/affluent; (2) aeration tank; (3) Sludge; (4) Chlorination tank; (5) Treated effluent.

Pseudomonas aeruginosa strains were recovered from all five stages of the HWTP studied, indicating that this species could be considered a good marker for studies aimed at evaluating the impact of wastewater treatment process on the prevalence of antibiotic resistance organisms and genes in the effluent. All stages of the station showed isolates exhibiting MDR pattern, in contrast, Slekovec *et al.* (2012), did not find resistant neither MDR isolates in HWTP before its release to the environment and in the sludge produced by the waste plant in France.

It is noteworthy to emphasize that 100% of the isolates of the chlorination step showed XDR profile. Using chemical disinfection to inactivate pathogens also plays an important role in controlling antibiotic-resistant bacteria in HWTPs. Studies on the effect of chlorination on antibioticresistant bacteria can be traced back to 1970s, where chlorination was shown to influence the proportion of multipleantibiotic-resistant bacteria in drinking water and wastewater (Grabow et al. 1973; Armstrong et al. 1982; Murray et al. 1984). On the other hand, in a previous study, we demonstrated the absence of microbial growth in the chlorination step. However, Ps. aeruginosa cells detected in the following stage of the treatment revealed the capacity of regrowth in chlorinated sewage effluent maybe due to the bacteriostatic effect of chlorine, which justifies the presence of viable cells in stage 5 (Santoro et al. 2012).

Among strains exhibiting β -lactamase encoding genes, bla_{TEM} , bla_{SHV} $bla_{CTX-M-1}$, bla_{KPC} , bla_{VIM} and bla_{SPM} were characterized randomly in 78% of the isolates. It is necessary to highlight that isolates carring bla_{TEM} , bla_{KPC} , bla_{VIM} and bla_{SPM} were found in the fifth and final stage of the hospital wastewater with potential to spread throughout the aquatic environment, thus enabling human exposure and transmission.

Quintera *et al.* (2005) reported the isolation of a *Pseu*domonas pseudoalcaligenes VIM-2 strain from hospital sewage. This discovery suggests that the ongoing spread of the $bla_{\rm VIM-2}$ is occurring simultaneously in several dimensions, as it can now be found in different environments and in several bacterial species. Another study revealed the presence of $bla_{\rm VIM-2}$ in two unrelated *Ps. aeruginosa* isolates from aquatic environments (Quintera and Peixe 2006). The $bla_{\rm SPM-1MBL}$ gene is the most prevalent in Brazil; its presence was evaluated in *Ps. aeruginosa* isolates from hospital sewage and surface-water samples, in order to obtain epidemiological data on the dissemination of this gene in environmental samples in southern Brazil (Fuentefria *et al.* 2009). The presence of

| 20 | 40 | 60 | 80 | 100 | | Profiles | n° of isolates | Date | Resistance pattern | Stage |
|----|-----------|--------|----|-----|------|----------|----------------|------|--------------------|------------------|
| | | | | | | 7 | 2 | 2010 | MDR(2) | 1(2) |
| | | _ | | | | 8 | 1 | 2010 | MDR(1) | 1(1) |
| | | | | | | 5 | 1 | 2008 | XDR(1) | 5(1) |
| | | ᅴ└ | | | | 2 | 2 | 2008 | MDR(2) | 2(1) |
| | _ | | | | ΪΪΪΪ | 6 | 4 | 2010 | MDR(2) | 1(1), 2(3) |
| | | | | | | 1 | 4 | 2008 | MDR(2), XDR(1) | 1, 3, 5(2) |
| | | | | | | 3 | 3 | 2008 | XDR(3) | 4(3) |
| | _ | | | | | 13 | 7 | 2010 | MDR(7) | 3(7) |
| | | | | | | 9 | 1 | 2010 | MDR(1) | 1(1) |
| Ч | L | \neg | | | | 11 | 1 | 2010 | MDR(1) | 2(1) |
| | | | | | | 12 | 1 | 2010 | MDR(1) | 2(1) |
| | | | | | | 10 | 1 | 2010 | MDR(1) | 2(1) |
| | | | | | | 18 | 1 | 2010 | | 5(1) |
| | _ | | | | | 19 | 1 | 2008 | MDR(1) | 2(1) |
| | \square | | | | | 17 | 4 | 2010 | MDR(2) | 5(4) |
| | | | | | | 20 | 3 | 2008 | MDR(2), XDR(1) | 3(1), 4(1), 5(1) |
| | 1 _ | | | | | 14 | 1 | 2010 | | 3(1) |
| Ц | | | 1 | | | 15 | 1 | 2010 | MDR(1) | 5(1) |
| | L | | | | | 16 | 1 | 2010 | MDR(1) | 5(1) |
| | | | | | | 4 | 1 | 2008 | MDR(1) | 5(1) |

Figure 2 Cluster analysis by enterobacterial repetitive intergenic consensus-PCR (ERIC-PCR) fingerprint (ERIC 2) of 41 *Pseudomonas aeruginosa* isolates. Clustering analysis was performed with aid of BIONUMERICS 6.6 (Applied Maths) and based on the Dice similarity coefficient and the unweighted pair group method with arithmetic mean.

| Table 3 | MLST | of isolates | ; representing | 20 <i>Ps</i> | eudomonas | aeruginos | a distinct | profiles | from I | hospital | wastewater | treatment | plants duri | ng 1 | 8-month |
|---------|------|-------------|----------------|--------------|-----------|-----------|------------|----------|--------|----------|------------|-----------|-------------|------|---------|
| period | | | | | | | | | | | | | | | |

| ERIC profile | | | ST | Allelic profile | | | | | | | | |
|--------------|----------------|----------------|------|-----------------|------|------|------|------|------|------|--|--|
| | No of isolates | Collected step | | acsA | aroE | guaA | mutL | nuoD | ppsA | trpE | | |
| 1 | 4 | 1, 3, 5 | 244 | 17 | 5 | 12 | 3 | 14 | 4 | 7 | | |
| 2 | 2 | 2 | 1853 | 16 | 22 | 11 | 5 | 4 | 4 | 10 | | |
| 3 | 3 | 4 | 244 | 17 | 5 | 12 | 3 | 14 | 4 | 7 | | |
| 4 | 1 | 5 | 244 | 17 | 5 | 12 | 3 | 14 | 4 | 7 | | |
| 5 | 1 | 5 | 1854 | 17 | 5 | 44 | 110 | 14 | 81 | 7 | | |
| 6 | 4 | 1, 2 | 1621 | 5 | 54 | 99 | 48 | 1 | 6 | 3 | | |
| 7 | 2 | 1 | 1855 | 5 | 54 | 44 | 48 | 14 | 81 | 3 | | |
| 8 | 1 | 1 | 1856 | 30 | 2 | 44 | 110 | 1 | 97 | 7 | | |
| 9 | 1 | 1 | 1857 | 28 | 54 | 44 | 110 | 4 | 79 | 3 | | |
| 10 | 1 | 2 | 1858 | 17 | 2 | 11 | 3 | 81 | 38 | 3 | | |
| 11 | 1 | 2 | 1859 | 17 | 5 | 36 | 7 | 27 | 4 | 5 | | |
| 12 | 1 | 2 | 1858 | 17 | 2 | 11 | 3 | 81 | 38 | 3 | | |
| 13 | 7 | 3 | 1859 | 17 | 5 | 36 | 7 | 27 | 4 | 5 | | |
| 14 | 1 | 3 | 381 | 11 | 20 | 1 | 65 | 4 | 4 | 10 | | |
| 15 | 1 | 5 | 1860 | 17 | 5 | 44 | 3 | 14 | 77 | 5 | | |
| 16 | 1 | 5 | 1861 | 17 | 20 | 49 | 110 | 11 | 79 | 10 | | |
| 17 | 4 | 5 | 1862 | 17 | 20 | 12 | 110 | 4 | 81 | 10 | | |
| 18 | 1 | 5 | 238 | 5 | 1 | 59 | 6 | 1 | 33 | 42 | | |
| 19 | 1 | 2 | 1853 | 16 | 22 | 11 | 5 | 4 | 4 | 10 | | |
| 20 | 3 | 3, 4, 5 | 595 | 17 | 5 | 12 | 5 | 14 | 4 | 7 | | |

ERIC, enterobacterial repetitive intergenic consensus; MLST, multilocus sequence typing; ST, sequence types.

pathogenic Enterobacteriaceae from a hospital effluent showed resistance to third-generation cephalosporins, ce-fepime, aminoglycosides, quinolones and a significant rate of carbapenem. Genes encoding ESBL (*bla*_{SHV}, *bla*_{TEM},

 $bla_{\text{CTX-M}}$ and bla_{GES}) were detected among 46 strains with reduced susceptibility to third-generation cephalosporins isolated from hospital sewage (Pitout *et al.* 2004; Coque *et al.* 2008; Naas *et al.* 2008).





The release of high concentrations of antibiotics and resistance genes in natural ecosystems is a recent event in evolutionary terms. However, these pollution can impact the structure and the activity of environmental microbial populations. Given that environmental micro-organisms are the original source of resistance genes acquired by human pathogens (Davies 1997). In water bodies, bacteria from different origins (human, animal, environmental) are able to mix, and resistance evolves as a consequence of promiscuous exchange and shuffling of genes and genetic vectors (Baquero *et al.* 2008).

The discharge of MDR bacteria including ESBL, KPC and MBL producers into an urban river is worrisome, as these isolates could persist in the environment and act as opportunistic pathogens and/or resistance reservoirs that could accelerate the evolution of antimicrobial resistance in the community (Kim and Aga 2007; Baquero *et al.* 2008; Martinez 2009).

The MLST analysis showed a high diversity, as reported in other previous studies (Gomila *et al.* 2013). ST244 is well distributed worldwide with 34 isolates in the MLST database being four isolated in Brazil (three from this study) and one clinical sample (Maâtallah *et al.* 2013; Bae *et al.* 2014; Chen *et al.* 2014;). Two strains belonging to ST595 deposited in the MLST database were

isolated in Brazil both from nonclinical samples (one from this study). It is noteworthy to point out that the chlorination step showed the greatest number of isolates of MDR and XDR strains belonging to CC244 where all the isolates from this step were grouped (Fig. 3). To our knowledge, this is the first MLST analysis of *Ps. aeruginosa* isolated from hospital effluent.

Following the rule of Maâtallah *et al.* (2013), 11 CC, sharing five or more identical alleles, were described with two major clones formed by five ST each. One of these CC (CC244) named by its central ST-244, included other closely related ST (ST-990, ST-993, ST-986 and ST-654). Two ST belonging to CC244 (ST-244 and ST-654), included several clinical strains showing reduced susceptibility against different classes of antibiotics such as ESBL, IMP-type metallo-beta-lactamase and KPC producing. We could not compare strains associated with ST-990, ST-986 and ST-654 also included in CC244, as these ST were deleted from the MLST database and no data associated with these strains was available.

The three strains classified as ST-595 were beta-lactamase producers and showed reduced susceptibility with one strain showing a KPC and metallo-beta-lactamase phenotype (MDR), one ESBL and metallo-beta-lactamase (XDR) and one ESBL producer (MDR). Other eight strains classified as ST-244 and also included in CC244 showed similar susceptibility patterns with several MDR and XDR strains KPC, ESBL and metallo-beta-lactamase producers. To date, CC244 has been detected among strains isolated from different sources such as clinical and environmental (Cholley *et al.* 2014). We first report, to our knowledge, the presence of strains belonging to CC244 from hospital effluent. These data suggest a possible turnover of strains sharing different antibiotic resistance mechanisms regardless of environment location.

The presence of MDR and XDR strains into CC244 suggests the pathogenic potential of this clone carrying strains with reduced susceptibility to several antimicrobial classes spreading through different sources from clinical to hospital effluent and the environment.

The spreading of MDR organisms and genes through the environment is worrying. Strategies to reduce this picture are required, including neutralization of antibiotics in wastewater and in the environment (Zwiener and Frimmel 2000; Ternes *et al.* 2003; Berglund *et al.* 2014).

Our results demonstrated the presence of isolates carrying important resistance mechanisms with potential to contaminate aquatic environments with increasing antibiotic resistance within the microbial community, leading to negative impacts on the environment and human health.

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Conflict of Interest

No conflict of interest declared.

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