

Nonvolant small mammals from a southwestern area of Brazilian Cerrado: diversity, habitat use, seasonality, and biogeography

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Introduction. The Cerrado is characterized by a mosaic of habitats ranging from open grasslands to dense arboreal savannas and woodlands, with gallery forests along watercourses. Extensive areas of Cerrado have been gradually deforested in the last 50 years, currently only 20% of the biome still retains its original vegetation, and only 2.5% is preserved as permanent protected areas. Scientific knowledge of the Cerrado fauna, such as composition, natural history, and geographic distribution of its species is very important to conservation strategies. Nevertheless, most of the Cerrado still remains poorly sampled and studied. Although an increasing number of works have been published for Cerrado nonvolant small mammals in recent years, there are few studies dealing with biogeographic patterns for the biome. The objective of this study was to investigate the diversity of nonvolant small mammals and to determine the habitat use of the species encountered at Emas National Park - ENP, one of the most important reserves of the Cerrado biome, located in a region yet poorly known.

Methodology. We sampled 28 different areas representing seven habitat types found in the ENP. Small mammals were surveyed during rainy and dry seasons with live and pitfall traps. The total effort was 10,664 trap-nights for live and 2,898 trap-nights for pitfall traps. The data analysis consisted of rarefaction curves, estimates of species richness, data on relative abundance, diversity and evenness indexes, as well as the faunal composition in each habitat type and in ENP in order to observe how the assemblage is characterized and distributed in the landscape to better understand the biogeography of this southwestern region in the Cerrado.

Results. We recorded 23 species of marsupials and small rodents in ENP. The grass mouse *Necromys lasiurus* was the most abundant, with the majority of species being rare. More species and individuals were encountered during the rainy season, with gallery forests and arboreal dense savannas representing the richest habitats. The assemblage was composed by two species groups: one inhabiting the forests and the other the open areas. Species composition overlaps with those of adjacent biomes and other Cerrado regions, but we found endemic and southwestern restricted species, characterizing an assemblage most similar to the Pantanal and Chaco biomes.

Discussion and conclusions: Previous studies in the Cerrado have shown a higher richness and abundance during the dry season. Our results could be explained by the higher abundance (reproduction period)

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and activity (abundance of resources) during the rainy season. The habitat selectivity results showed the great importance of canopy and humidity on the distribution of small mammals, with gallery forest being not the richest habitat in the region. The high richness also found in the savannas, although less stratified vertically, may be a result of horizontal heterogeneity. Other biomes adjoin the Cerrado and influence its faunal composition; compared to the Amazon and Atlantic Rainforest, the fauna of the present study showed greater similarity with the Amazon, contrary to prior studies. This pattern is mainly related to the presence of four taxa that occur in both the Amazon and Cerrado, but are replaced by sister groups in the Atlantic Rainforest. This study highlights the importance of this protected area, as it harbors a high richness of nonvolant small mammals, including endemic species and also species with restricted distributional ranges within the Cerrado, together with a unique faunal composition, characteristic of the southwestern portion of this biome.

Key words: Assemblage, Hotspot, Marsupials, National Parks, Rodents, South America.

Resumen

El objetivo de este estudio fue investigar la diversidad de pequeños mamíferos no voladores y determinar su uso de hábitat en el Parque Nacional Emas – PNE, una de las reservas más importantes del bioma Cerrado, ubicado en una región aún poco conocida. Registramos la presencia de 23 especies de marsupiales y pequeños roedores en el PNE, 20 de las cuales fueron encontradas en siete tipos distintos de hábitats mediante el uso de trampas de caída (2,898 noches-trampa, 9 especies) y trampas de captura viva (10,664 noches-trampa, 18 especies). En general, fueron encontradas más especies durante la época lluviosa (20) que durante la época seca (12). El hábitat analices mostró claramente un ensamblaje compuesto por especies habitando diferentes tipos de hábitat, obteniéndose dos agrupamientos principales: bosques y áreas abiertas; pocas especies fueran observadas en ambos tipos de hábitat, fisonómicamente diferentes. Desde el punto de vista biogeográfico, la composición de especies registrada en el PNE se superpone con la de los biomas adyacentes cuando se consideran las especies generalistas y con amplia distribución. Sin embargo, encontramos cuatro especies endémicas del Cerrado, tres de ellas exclusivas de las formaciones abiertas del parque. La mayor similitud en cuanto a composición se dio con la fauna del Pantanal, seguida por la del Chaco, la de la Caatinga y la del Amazonas. El presente estudio resalta la importancia de esta área protegida, dado que contiene una alta riqueza de pequeños mamíferos no voladores, incluyendo tanto especies endémicas como especies con distribuciones restringidas dentro del Cerrado y por su composición faunística única característica de la porción sudoeste de este bioma.

Palabras Clave: América del Sur, Ensamblaje, “Hotspot”, Marsupiales, Parques Nacionales, Roedores.

Introduction

The Cerrado, the second largest South American biome, with more than 2 million km² mainly located in the highland plateaus of Central Brazil (Eiten 1994), is characterized by a mosaic of habitats ranging from open grasslands to dense arboreal savannas and woodlands, with gallery forests along watercourses (Ribeiro and Walter 1998). Extensive areas of Cerrado have been gradually deforested in the last 50 years for the installation of mechanized agriculture and extensive cattle raising programs (Klink and Machado

2005). Currently only 20% of the biome still retains its original vegetation (Myers *et al.* 2000), and 2.5% is preserved as permanent federally protected areas (Klink and Machado 2005). Scientific knowledge of the Cerrado fauna, such as composition, natural history, and geographic distribution of its species is very important to conservation strategies (Silva *et al.* 2006). Nevertheless, most of the Cerrado still remains poorly sampled and studied (Werneck 2011).

The mammalian fauna of the Cerrado is rich, represented by at least 227 species in its core area (Carmignotto *et al.* 2012) and by 251 species including the transitional areas (Paglia *et al.* 2012). However, these species are not distributed homogeneously throughout the biome, and often are distributed in a restricted and/or heterogeneous manner, and this is especially true for small mammals (Alho 2005). Studies in the Cerrado have suggested a patchy distribution for the nonvolant small mammals which are associated with the distinct habitats in the region (*e. g.* Mares *et al.* 1986; Lacher and Alho 2001; Santos-Filho *et al.* 2012), with distinct species assemblages found in different areas (Carmignotto *et al.* 2012). Also, the relative abundance of species varies among localities, and some species can be found in high numbers at a given site, but be rare or even absent in others (*i. e.* Marinho-Filho *et al.* 1994). Usually, most species tend to be locally rare and few are very abundant within all or most assemblages (Vieira and Palma 2005).

Although an increasing number of works have been published for Cerrado nonvolant small mammals in recent years, there are few studies dealing with biogeographic patterns for the biome (see Werneck 2011). Since the effort is not homogeneous across this vast area, apparent global trends can be biased by the varying extent of information among different portions of the Cerrado, underestimating or enhancing the importance of some abiotic and/or environmental factors associated with the distribution of these mammals (see Vieira *et al.* 2008).

The aim of this study was to investigate the diversity of the nonvolant small mammals and to determine the habitat use of the species surveyed at Emas National Park (ENP), one of the most important protected areas of Cerrado, located in a poorly known region with the purpose of increasing the knowledge of this area, thus contributing to a better understanding of global biogeographic patterns for the biome.

We have published a mammal checklist from Emas National Park (Rodrigues *et al.*, 2002), but in the present work we have re-identified the nonvolant small mammal species based on new genera and species described and/or resurrected from synonyms in recent years (Langguth and Bonvicino 2002; Bonvicino 2003; Voss *et al.* 2005; Carmignotto and Monfort 2006; Weksler *et al.* 2006; Bezerra and Oliveira 2010) and also we have specifically analyzed the ecological data in relation to this mammalian group, such as their richness, abundance and composition in the distinct habitats of the ENP and also during the different seasons of survey.

Study area.

Emas National Park (ENP) has an area of 132,642 ha (ICMBio 2013) and is located in extreme southwestern Goiás state, near the limits with Mato Grosso and Mato Grosso do Sul states (-18°15' S, -52°53' W; Fig. 1). The Central Brazilian plateau, with the higher areas at 800 meters above sea level, concentrates the headwaters of the most important

hydrographic basins in the continent and specifically in and around ENP are headwaters of rivers that form the Amazon (Araguaia River) and Paraná (tributary of the Paranaíba River) basins, as well as the Taquarí River, a tributary of the Paraguay basin, in the Pantanal complex. This makes this region biogeographically important, due to the convergence and coexistence of elements from distinct ecosystems, such as the Pantanal, the Atlantic Rainforest, and the Amazon. Elevation in the ENP ranges from 650 – 1,000 m, part of the “Chapadões de Goiás” (Radambrasil 1983).

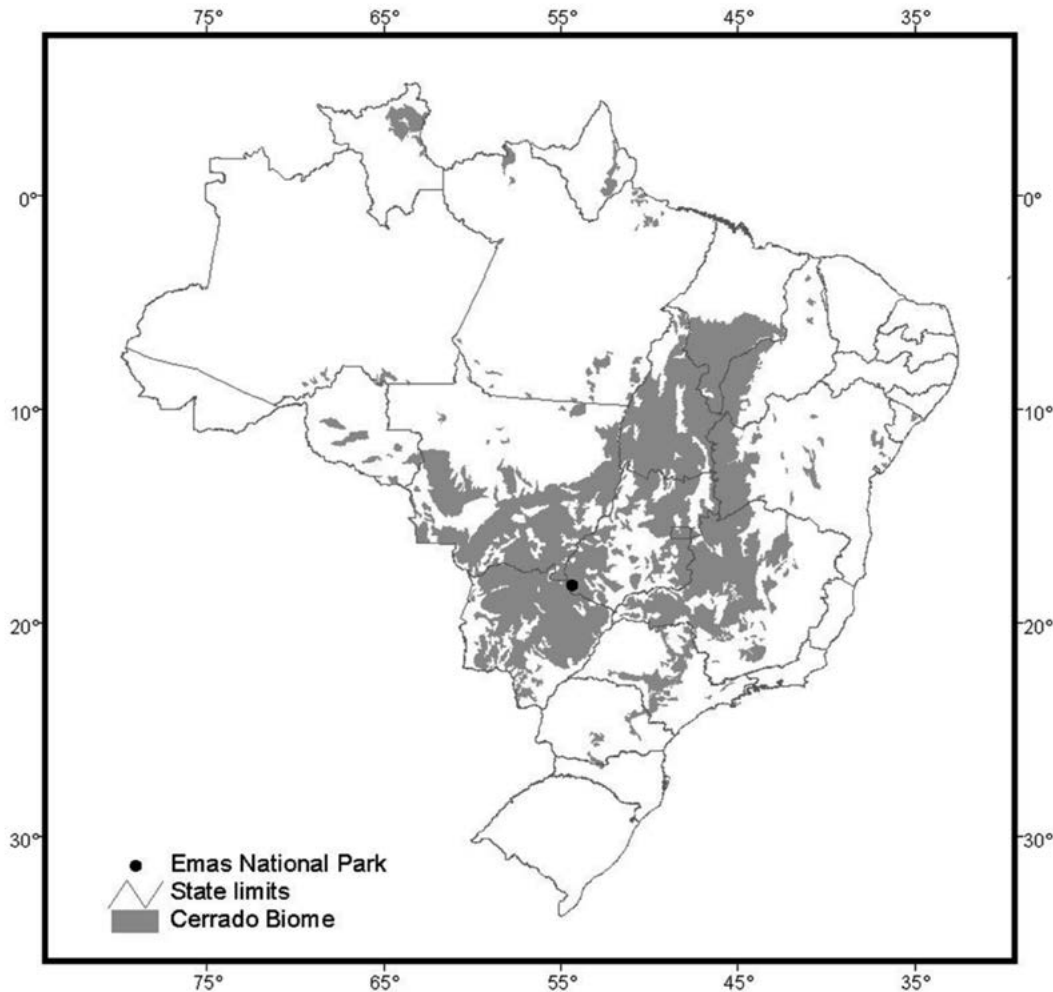


Figure 1. The location of Emas National Park (ENP) in Central Brazil, including the limits of the Cerrado biome based on IBGE (1993).

The climate of ENP is classified as Tropical, type AW according to Köopen’s classification, with two well defined seasons: the dry season, from May to September, and rainy season, from October to April, with total precipitation varying from 1,200 to 3,000 mm per year (D’Angiolella 2004). The temperature varies from -3 to 40 °C, with the mean annual temperature around 20 – 22 °C. (D’Angiolella 2004).

Landscape of ENP can be described as follows: (1) flooded (MGI) and (2) unflooded (MG) evergreen gallery forests along the watercourses; (3) flooded grasslands (CU) characterized by open and homogeneous vegetation composed of grasses, some with the presence of palms forming lines, the so called “veredas”; unflooded grasslands, which can be divided in different types according with the density of the scrub; (4) “campo limpo” (CL), open grassland; (5) “campo sujo” (CS), open grassland with some trees and

shrubs; (6) “campo cerrado” (CC), shrubby vegetation with a ground cover of grasses; and (7) the arboreal dense savannas “cerrado *sensu stricto*” (CE), woodland with an open canopy and some grass covering the ground (Ribeiro and Walter 1998).

Material and Methods

Study area and sampling procedures

We sampled 28 different areas in ENP that represent the seven habitat types found in the study site described above (Table 1). Nonvolant small mammals were surveyed during four different periods: two during the rainy season (November 3 - 14, 1998 and February 11 - 25, 1999) and two during the dry season (May 3 - 15 and July 26 - August 6, 1999).

We applied two methods of capture, live and pitfall traps. These methods were used because they have been shown to be complementary, capturing different species and individuals (e.g., Voss and Emmons 1996). The live traps were placed 10 meters apart along linear transects, using both Sherman (7.5 × 8.5 × 23 cm) and Young (19.5 × 20 × 32 cm) traps (2/3 Sherman and 1/3 Young traps) to increase the probability of capture of species of different body sizes, comprising transects around 400 meters long (40 live traps). Transects were at least 1 km apart from each other. The bait was a mixture of peanut butter, sardine, banana, and ground maize, adding a little piece of corn in the Young traps. The total effort for live traps was 10,664 trap-nights (see Table 1 for the capture effort in each habitat type).

Three pitfall trap capture stations were installed in five of the seven habitat types (Table 1). Each pitfall capture station was composed of six buckets of 35 liters, spaced three meters from each other disposed in the form of the letter “T”, for a total effort of 2,898 trap-nights (Table 1).

Table 1. Habitats surveyed and the effort of live and pitfall traps in each habitat type during each season and the study period. In parentheses the number of areas surveyed per habitat.

Habitat	Live Trap Effort			Pitfall Effort		
	Rainy	Dry	Total	Rainy	Dry	Total
1- Gallery forest (5)	1,235	1,152	2,387	306	180	486
2- Flooded gallery forest (1)	479	199	678	-	-	-
3- “Campo limpo” (1)	238	-	238	-	-	-
4- “Campo sujo” (5)	998	1,610	2,608	486	954	1,440
5- “Campo cerrado” (3)	777	921	1,698	144	144	288
6- Flooded grassland (5)	706	120	826	126	-	126
7- Arboreal dense savanna (8)	1,904	325	2,229	468	90	558
Total (28)	6,337	4,327	10,664	1,530	1,368	2,898

Nomenclature and classification follow Wilson and Reeder (2005) for the rodents, and Gardner (2008) and Voss and Jansa (2009) for the marsupial species. We also followed Bonvicino *et al.* (1996) for the genus *Nectomys*, Weksler and Bonvicino (2005) for *Oligoryzomys*, Weksler *et al.* (2006) for *Hylaeamys*, Almeida *et al.* (2007) for the genus *Calomys*, and Percequillo *et al.* (2008) for *Cerradomys*. External and craniodental

morphology as well as karyotypic data were used to confirm identification of species.

Specimens were collected under a permit from IBAMA (license number 102/98 – DIFAS). The specimens are deposited in the Museu Nacional, Universidade Federal do Rio de Janeiro (MN), the Departamento de Zoologia, Universidade de Brasília (UNB), and the Museu de Zoologia, Universidade de São Paulo (MZUSP; Appendix 1).

Data analysis

The effectiveness of the survey was analyzed through a sample-based rarefaction curve with the number of taxa plotted against the accumulated number of individuals as recommended by Gotelli and Colwell (2001). The total capture effort (by live and pitfall traps), and number of individuals and species sampled in each trap transect, were used to obtain the taxon sampling curve. Additionally, a sampling curve for each season was made to compare the richness among these survey periods. In both cases, the rarefaction curve was preferred in relation to the accumulation curve because it represents the statistical approach of the corresponding accumulation curve and produces a smooth curve that is better compared with results from other studies (Gotelli and Colwell 2001).

Rarefaction was also used to compute species richness estimates with the software ESTIMATE S version 8.0.0 (Colwell 2004). The expected richness for the study area was estimated using 100 random samples without replacement based on the second-order Jackknife estimator because it performs well in assemblages with moderate to low evenness indices, as is the case in the present study (Brose *et al.* 2003). Mean and standard deviation for the observed and estimated richness were also computed.

Relative abundances of species were estimated based on the total number of individuals sampled per species in comparison with the total number of specimens surveyed during the study period (Magurran 2004). The relative abundance or capture frequency of each species in the assemblage was graphically represented in a Whitaker plot (Whitaker 1972). The Shannon-Wiener diversity index (H' - natural log) corrected by number of species ($D = e^{H'}$), known as Hill's diversity number (N_1), was used to estimate the diversity of the assemblage sampled. This transformed diversity measure is meaningful because it permits direct comparisons and is sensitive to the number of rare species in the communities (Ludwig and Reynolds 1988; Jost 2010). Pielou's J was used as a measure of evenness, or the relative contribution of each species for the assemblage diversity, since it is one of the most well-behaved evenness measures (Magurran 2004; Jost 2010).

The concept of additive partitioning (Veech *et al.* 2002) was used to obtain the relative contribution of each distinct level of diversity, the " α " and " β " diversity, for the total diversity found in the study area, the " γ " diversity or total richness (Whitaker 1972).

The " α " diversity was considered as the mean number of species found in each habitat surveyed, and the " β " diversity as the mean species richness found among habitats. One of the advantages of using the additive and not the multiplicative concept of Whitaker (1972) is that " α " and " β " diversity are estimated in the same unit of measure, can be treated as proportions of the total diversity (" γ " diversity), and can be estimated on multiple spatial scales, enabling inferences about the biological meaning of these measures (Veech *et al.* 2002; Pardini and Umetsu 2006). Because capture effort differed for each habitat surveyed, the capture success (%) - calculated as the total number of

individuals surveyed in each habitat divided by the total capture effort in each habitat - was used as a comparative measure.

A Detrended Correspondence Analysis (DCA), known as DECORANA (Hill and Gauch 1980), was used to evaluate the relationship among habitats based on the composition and abundance of nonvolant mammal species (represented by the capture success in each habitat type- %). This method was chosen due to the simultaneously visual ordination of both habitat types and species, which facilitates interpretation of the results, and due to the nonlinear species relationships found in this study (Urban 2000). We used the software PC-ORD for Windows 4.25 (McCune and Mefford 1999) with the rescale axes option and 26 as the number of segments (Palmer 2004).

Data on the occurrence of species in the Brazilian and Central South America biomes were based on Paglia *et al.* (2012) and Carmignotto *et al.* (2012), respectively.

Results

We recorded 23 species of nonvolant small mammals in ENP (Table 2). The mouse opossum *Thylamys velutinus* (Wagner, 1842) was found dead along a rail track of the Park, and the water opossum *Chironectes minimus* (Zimmermann, 1780) and the gray four-eyed opossum *Philander opossum* (Linnaeus, 1758) were recorded only by direct observations and were not included in the analyses. Twenty species were captured either in pitfalls (9 species), or in live traps (18 species). The gracile mouse opossum *Cryptonanus* sp. and the short-tailed opossum *Monodelphis kunsii* Pine, 1975 were only captured by pitfalls, whereas 11 species were captured only by live traps (Table 2). The live traps had a total trapping success of 5.5% and the pitfalls of 0.7 %. The live traps had a greater number of captures during the dry season (4.1% rainy – 7.6 % dry). On the other hand, the pitfalls had more captures during the rainy season (1.1 % rainy - 0.3% dry).

The grass mouse *Necomys lasiurus* (Lund, 1841) was the most abundant species, with seven species considered as common (Fig. 2) and the rice rat *Cerradomys marinus* (Bonvicino, 2003) and the grass mouse *Calomys expulsus* (Lund, 1841) as uncommon species, with the majority of species surveyed considered rare in the assemblage (Fig. 2). When we corrected the data due to the different capture effort made among the two seasons and seven habitat types (capture success), we found that only the grass mouse *Necomys lasiurus* had an increase in the number of individuals sampled during the dry season (Table 2).

Considering the results obtained by the rarefaction curves, the estimated richness for the study area was 24 species. So, the sample realized in the present study represented around 83.3% of the species richness of the region (Fig. 3). Indeed, if the three additional species are considered, this survey in the ENP sampled 96 % of the nonvolant small mammal assemblage expected to occur in the Park. These results are supported by the stabilization of the number of species after around the survey of 200 individuals or 8,000 trap-nights (Fig. 3). In relation to seasonality, the rarefaction curves were also near to asymptotes, especially for the survey during the rainy season (Fig. 4). Considering the same capture effort (similar number of individuals surveyed) we captured more species during the rainy than in the dry season (19 rainy - 12 dry, Fig. 4). Observing the capture effort applied in each of the seven distinct habitats, the majority of the habitat types

presented curves near to stabilization, except the arboreal dense savanna or “cerrado *sensu stricto* = CE” (Fig. 5). Although this habitat showed the highest richness found for a single habitat, we would expect more species based on rarefaction curves. A mean capture effort of around 1,200 trap-nights (or 50 individuals) in each habitat was necessary for species richness to begin to reach an asymptote (Fig. 5). Excluding the “campo limpo” due to low capture effort and individuals surveyed, we can observe that the richest habitats (considering the red line – similar capture effort) were the gallery forests (MG) and the arboreal dense savannas (CE), with the flooded habitats (MGI and CU) and the dry grasslands (CS and CC) presenting lower richness (Fig. 5).

Species	Live Trap	Pitfall	Total Period	Rainy Season		Dry Season	
	N	N	N (RA %)	N	CS (%)	N	CS (%)
ORDER DIDELPHIMORPHIA							
Family Didelphidae							
<i>Chironectes minimus</i> (Zimmermann, 1780)*	-	-	-	-	-	-	-
<i>Cryptonanus</i> sp.**	-	04	04 (1.1)	04	0.4	-	-
<i>Didelphis albiventris</i> Lund, 1840	23	-	23 (6.4)	18	0.9	05	0.5
<i>Gracilinanus agilis</i> (Burmeister, 1854)	03	-	03 (0.8)	02	0.2	01	0.1
<i>Lutreolina crassicaudata</i> (Desmarest, 1804)	04	-	04 (1.1)	03	0.2	01	0.3
<i>Marmosa murina</i> (Linnaeus, 1758)	17	-	17 (4.7)	15	1.0	02	0.2
<i>Monodelphis domestica</i> (Wagner, 1842)	02	-	02 (0.6)	02	0.3	-	-
<i>Monodelphis kuni</i> Pine, 1975	-	01	01 (0.3)	01	0.1	-	-
<i>Philander opossum</i> (Linnaeus, 1758)*	-	-	-	-	-	-	-
<i>Thylamys velutinus</i> (Wagner, 1842)***	-	-	-	-	-	-	-
ORDER RODENTIA							
Family Cricetidae							
<i>Calomys expulsus</i> (Lund, 1841)	05	02	07 (1.9)	07	0.1	-	-
<i>Calomys tener</i> (Winge, 1887)	02	02	04 (1.1)	04	0.2	-	-
<i>Cerradomys marinus</i> (Bonvicino, 2003)	11	-	11 (3.0)	09	0.4	02	0.5
<i>Cerradomys scotti</i> (Langguth and Bonvicino, 2002)	18	01	19 (5.2)	15	1.7	04	0.6
<i>Hylaeamys megacephalus</i> (Fischer, 1814)	26	01	27 (7.5)	20	0.5	07	0.1
<i>Kunsia tomentosus</i> (Lichstentein, 1830)	05	-	05 (1.4)	04	0.3	01	0.1
<i>Necomys lasiurus</i> (Lund, 1841)	150	06	156 (43.1)	49	1.4	107	7.0
<i>Necomys rattus</i> (Pelzeln, 1883)	05	-	05 (1.4)	05	0.2	-	-
<i>Oecomys bicolor</i> (Tomes, 1860)	16	01	17 (4.7)	14	0.7	03	0.2
<i>Oligoryzomys fornesi</i> (Massoia, 1973)	02	01	03 (0.8)	03	0.1	-	-
<i>Oxymycterus delator</i> Thomas, 1903	21	-	21 (5.8)	18	0.5	03	0.2
Family Echimyidae							
<i>Clyomys laticeps</i> (Thomas, 1909)	28	-	28 (7.7)	19	0.5	09	0.4
<i>Proechimys longicaudatus</i> (Rengger, 1830)	05	-	05 (1.4)	05	0.4	-	-
Total			362 100%	217 60%		145 40%	

Table 2. Species of nonvolant small mammals encountered at Emas National Park (ENP) and the number of species and individuals sampled by each method used. Number of individuals (N) and relative abundance (RA %) in the total period of the study, and the capture success (CS) along the rainy and the dry season. The capture success was calculated using only habitats where each species was captured.

* These species were recorded only by direct observation in the Park; *C. minimus* in an area of “campo limpo” and *P. opossum* of “campo úmido”.
 ** This taxon is distinct from the five valid species of this genus (*sensu* Voss *et al.*, 2005), representing a new taxon yet undescribed in the literature.
 *** One individual was found dead along a rail track of the Park near areas of “campo sujo” (grassy and shrubby vegetation).

The transformed Shannon index (D) was 8.76 and the Pielou’s evenness index (J) 0.72 over the total period of study. However, we also noted a variation of these values during

the different seasons when considering the same number of individuals surveyed (Fig. 4), with the rainy season showing greater values for diversity and evenness ($D = 12.68$, $J = 0.86$ rainy and $D = 3.16$, $J = 0.46$ dry).

The results obtained by the additive partition analysis, with a mean of seven species encountered in each habitat type (“ α ” diversity) and of 13 species among habitat types (“ β ” diversity), demonstrate the selectivity of the species for the distinct habitats sampled.

This resulted in a mosaic pattern of species richness distribution in the study area, with 35% of the diversity being distributed in each habitat and 65% among the habitat types.

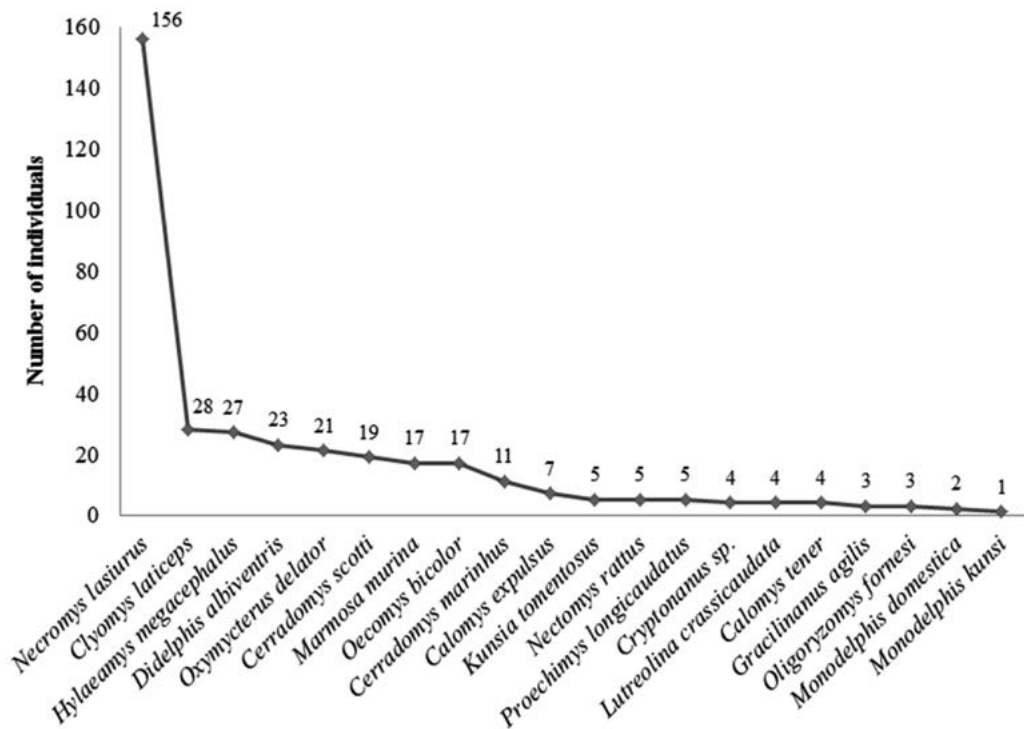


Figure 2. Relative abundance of the twenty nonvolant small mammal species encountered in ENP.

In the study area, the gallery forests (flooded and unflooded) harbored 10 species, whereas the number of species that occurred in the open dry formations (“campo limpo”, “campo sujo”, “campo cerrado” and arboreal dense savanna) totaled 13. The flooded grassland was one of the poorest habitats, composed of four species (*L. crassicaudata*, *N. rattus*, *O. fornesi* and *O. delator*) which, however, were quite distinctive from the species abundance and composition of the other habitats, forming an independent group (see Fig. 6). We found 7 species (35 %) inhabiting only the forests and 10 species (50 %) restricted to the open formations, with *D. albiventris*, *L. crassicaudata* and *N. rattus* occurring in both habitat types (15 %). Some species were restricted to a single habitat such as the marsupials *G. agilis* and *M. kunsii*, the rice mouse *H. megacephalus* and the spiny rat *P. longicaudatus*, which were found only at the unflooded gallery forest; *K. tomentosus* was restricted to areas of “campo sujo” and *M. domestica* to arboreal dense savanna habitats.

Among the open areas, the flooded areas were grouped apart from the dry ones due to the co-occurrence of species with gallery forests, such as *Lutreolina crassicaudata* and *Nectomys rattus* (Fig. 6). Areas of “campo limpo”, represented by only one species,

N. lasiurus, joined areas of “campo sujo” and arboreal dense savanna which presented similar species richness (9 and 10 species), abundance and composition (Fig. 6). In the forested habitats, the assemblage of the flooded gallery forest was poorer, represented by five species, with three of them very abundant in this habitat (*M. murina*, *C. marinhus* and *O. bicolor*) leading to a separate grouping (Fig. 6). The forests and open formations showed distinct distributions along Axis 1 of the DCA (eigenvalue = 0.983), whereas the flooded and unflooded gallery forests can be distinguished in Figure 6 by values from Axis 2 (eigenvalue = 0.521).

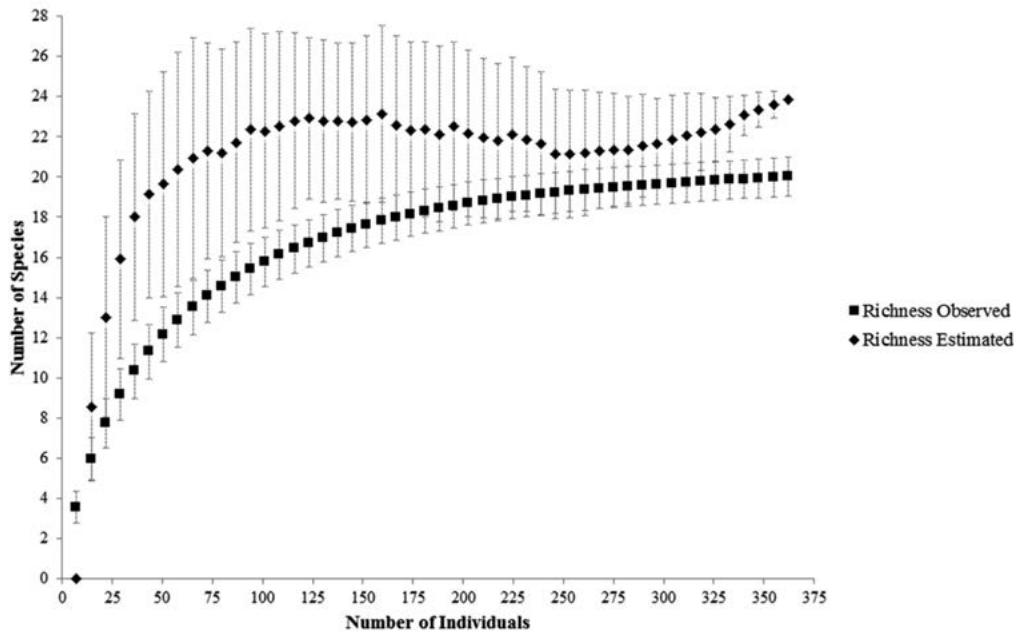


Figure 3. Sample-based rarefaction curves with the number of species plotted against the accumulated number of individuals during the study period, showing the observed and estimated richness (based on second order “Jackknife”), including the mean and standard deviation.

Analyzing the geographic distribution of each species surveyed in ENP (see Table 3), these results showed that most of the species surveyed (81 %) occur in more than one biomes, with four species (*Cryptonanus* sp., *Thylamys velutinus*, *Cerradomys marinhus* and *Cerradomys scotti*; 19 %) being endemic to the Cerrado. The greatest faunal overlap was between the Cerrado and the Pantanal, with 16 species in common (69 %). The next biome with 48 % species in common was the Chaco, followed by the Caatinga (43 %) and the Amazon (39 %; Table 3).

Discussion

The live trap success rates obtained in this study was similar (2-5 %) to those found in other studies in the Cerrado where different habitats were surveyed (e. g., Marinho-Filho *et al.* 1994; Lacher and Alho 2001; Cáceres *et al.* 2011; Carmignotto and Aires 2011; Bonvicino *et al.* 2012; Santos-Filho *et al.* 2012; Owen 2013). However, some studies dealing with only one type of habitat (e. g. Ribeiro and Marinho-Filho 2005; Becker *et al.* 2007) or situated in transitional areas (*i. e.* Bezerra *et al.* 2009) presented higher capture success rates (7-23 %). The pitfall trap success (around 1 %) was lower than that of live traps, similar to results found by Bezerra *et al.* (2009) and Carmignotto and Aires (2011) but very distinct from the high pitfall success (7.5 %) presented by Cáceres *et al.* (2011). These later authors have suggested that the bucket size is the variable that could explain these results, since the majority of studies used smaller buckets (20-35 liters) instead of

larger ones (60-108 liters) that are more effective for trapping small mammals.

In the present study we observed a seasonal variation in the number of species and individuals captured by the live and pitfall traps, with more species and individuals being captured during the rainy season. Nevertheless, the only species that had an increase in the number of captures during the dry season was the grass mouse *Necromys lasiurus*.

This species was responsible for the greater live trap success during the dry season. Considering that it is a seed eating species, the increase in this species population could be reflective of seed production by the grassland plant species, as was shown in other areas of Cerrado (*i. e.* Vieira *et al.* 2005). This result diverges from that of Becker *et al.* (2007) that found a dominance of *N. lasiurus* in both seasons in ENP, but not differing in numbers between the rainy (January-February) and dry-wet (October) season. This difference is probably due to the absence of trapping in the dry season (May-September) by those authors.

The studies so far have shown a higher richness and abundance during the dry season (*i. e.* Ribeiro and Marinho-Filho 2005). The results obtained in the present study, with more species being surveyed during the rainy rather than in the dry season, could be explained by the higher abundance and/or rhythm of activity displayed by the species during the rainy season, as was shown by the higher rate of capture in the pitfall traps, which did not have a bait to attract the individuals. It is possible that the higher trap success is related mainly with the reproductive season and the resource availability, which regulate the degree of activity of these mammals. Thus, the activity of the majority of the species surveyed in the present study could be influenced by the higher availability of resources during the rainy season, such as fruits (*i. e.* Oliveira 1998) and insects (*i. e.* Pinheiro *et al.* 2002) and by the increase in the density of individuals through reproduction (*i. e.* Francisco *et al.* 1995). During the dry season we believe the mammals were less active, therefore traveled shorter distances and were captured mainly by the live traps because of the attraction of the bait, as there are few resources available in this season, comparatively. The exceptions were some rodent species that feed on grasses, such as the grass mouse *Necromys lasiurus*. However, as the small mammal species of the Cerrado show great variation in population density, time of reproduction and rhythm of activity among different localities (*e. g.* Santos-Filho *et al.* 2012; Owen 2013), only long-term studies could reveal the relationships between resource abundance, density and reproductive activity in this area.

The Emas National Park nonvolant small mammal assemblage is very rich (23 species).

This richness is high when compared with other areas of Cerrado (Mares *et al.* 1986 – 25; Marinho-Filho *et al.* 1994 – 4 to 17; Lacher and Alho 2001 – 19; Bezerra *et al.* 2009 – 13; Rocha *et al.* 2011 – 22; Carmignotto and Aires 2011 – 24; Bonvicino *et al.* 2012 – 19; Santos-Filho *et al.* 2012 – 21) and is comparable to the numbers found in Amazonian, Atlantic Rainforest and Caatinga (Chapada Diamantina) nonvolant small mammal assemblages (*e. g.* Voss and Emmons 1996 – 15 to 33; Vivo and Gregorin 2001 – 30; Oliveira and Pessôa 2005 – 25; Pardini and Umetsu 2006 – 23; De La Sancha 2014 – 19). This result probably was achieved by the great capture effort of live traps and number of habitats surveyed, which were higher than in the previous studies (see Marinho-Filho *et al.* 1994), and also because the use of different methods to survey the small mammals (live – Sherman and Young traps, and pitfall traps), reflecting the great variability of ecological aspects, such as different body size, diet, locomotion patterns

and behavior of the small mammal species (see Voss and Emmons 1996).

The diversity obtained can also be considered high ($D = 8.76$) when compared with other areas surveyed in the Cerrado, but not the evenness index ($J = 0.72$). In the study of Marinho-Filho *et al.* (1994), the diversity obtained along 10 different localities varied from $D = 1.77$ to 13.60 and the evenness index from 0.41 to 0.94. Carmignotto and Aires (2011) obtained a diversity between $D = 5.31$ to 11.82 and an evenness between $J = 0.76$ to 0.85 for three localities in the northern portion of the Cerrado. The low evenness index obtained in the present study characterizes an assemblage with species with high abundance variation. The considerable increase in the number of individuals of the grass mouse *Necromys lasiurus* and the total absence of some species during the dry season greatly contributed to this pattern. During the rainy season the relative abundance of the species was more even and the diversity index increased. This highlights the importance of seasonal variation among the small mammal assemblages in the Cerrado, altering the number of individuals and species, consequently modifying the assemblage structure along the year. The assemblage sampled showed a dominant species (*Necromys lasiurus*), some species of intermediate abundance (7) and many rare species represented by few individuals surveyed (12). In general, the Cerrado nonvolant small mammal assemblages are characterized by this pattern of relative abundances (e. g. Lacher and Alho 2001; Vieira and Palma 2005; Bezerra *et al.* 2009; Carmignotto and Aires 2011).

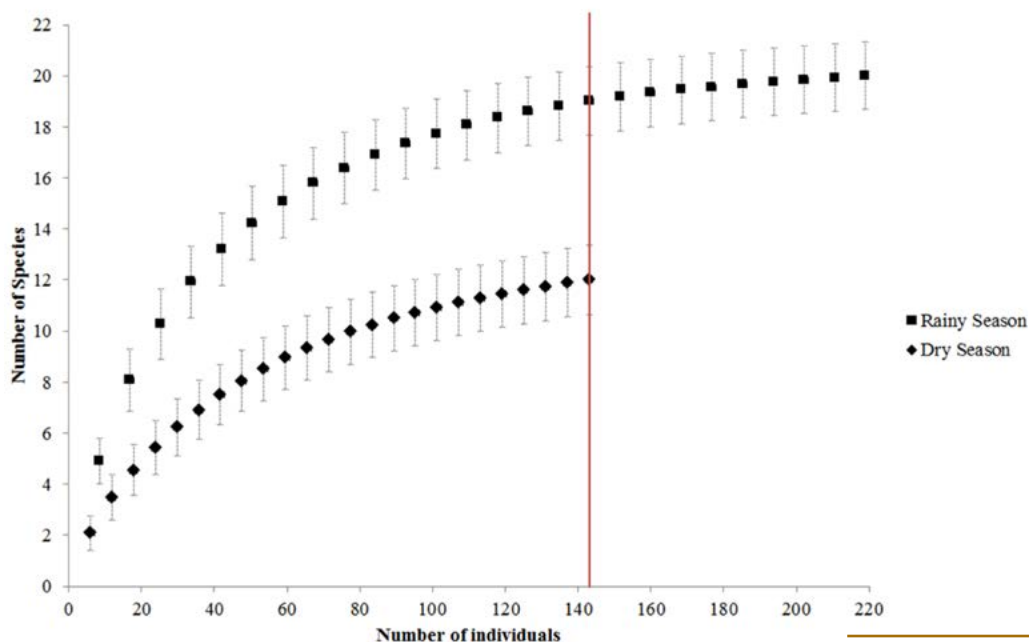


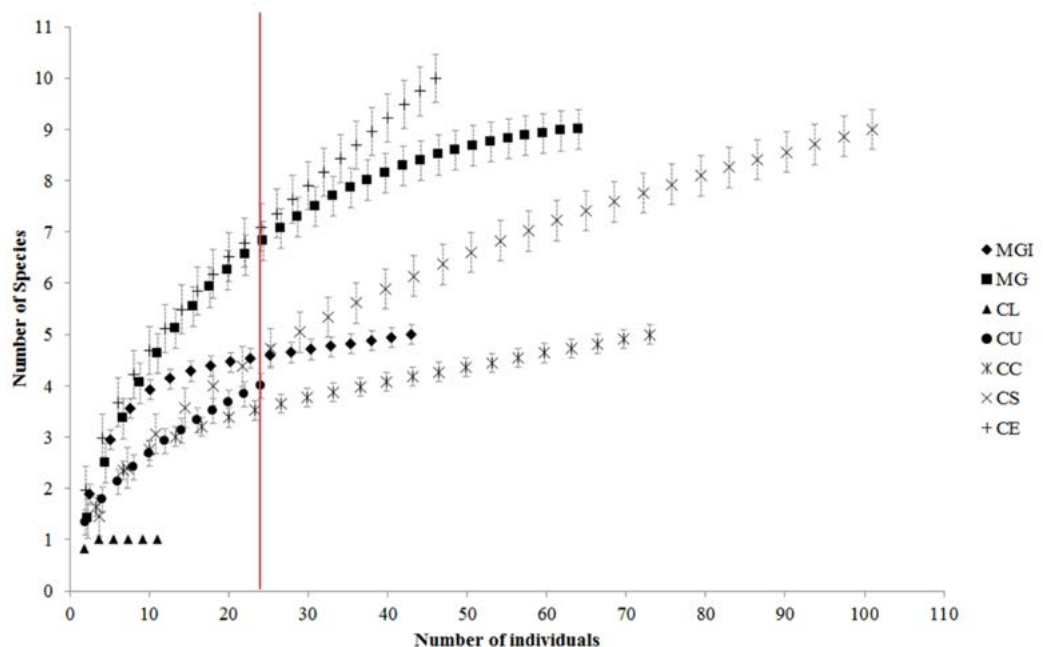
Figure 4. Sample-based rarefaction curves with the number of species plotted against the accumulated number of individuals during each season (rainy and dry), including the mean and standard deviation. The red line indicates the common minimum number of individuals (similar effort), permitting richness comparison among seasons.

The DCA and additive partition analysis clearly showed that the assemblage is composed by species that inhabits different types of habitats, being one reason for the high richness found, since they do not live in the same area, decreasing the probability of competition (see Lacher *et al.* 1989). Among the 20 species surveyed seven were restricted to forested areas and 10 to the open habitats, resulting in two principal groupings (the forests and the open areas inhabitants), a common pattern found in other nonvolant small mammal assemblages of the Cerrado (see Mares *et al.* 1986; Alho 2005; Carmignotto *et al.* 2012). Indeed, there was a finer segregation even between these environments, with the forest

species showing different preferences among the flooded and unflooded gallery forests, and the open areas inhabitants differing in species composition and capture rate among the dry and wet grasslands. In addition, these data showed the great importance of the canopy and humidity on the distribution of the nonvolant small mammal species of the Cerrado. In a habitat analysis by genus level of Cerrado small mammals, Vieira and Palma (2005) also highlighted the importance of the arboreal cover for their distribution, but the humid areas were not as relevant for the occurrence of genera as they are for some species in the present study. Carmignotto and Aires (2011) and Santos-Filho *et al.* (2012) also found distinct species composition and abundance in different kinds of gallery forests. These results indicate that the use of detailed habitat category, such as flooded and unflooded gallery forests, will help to find finer associations between the habitat structure and the small mammal species.

The great habitat selectivity by the nonvolant small mammals from the Cerrado is reflected in the distribution of the diversity in the communities. Higher richness can be found among distinct habitats, so the “ β ” diversity is more important (around 65-70 %) than the “ α ” diversity (around 30-35 %) when considering the richness present in the a study area, a common pattern found in this biome (e. g. Hannibal and Cáceres 2010; Carmignotto and Aires 2011). On the other hand, for the Atlantic Rainforest for example, the “ α ” diversity or local diversity (number of species in a given habitat or site) represented more than 50% of the richness in two localities surveyed (Pardini and Umetsu 2006).

Figure 5. Sample-based rarefaction curves with the number of species plotted against the accumulated number of individuals for each habitat type surveyed, including the mean and standard deviation. The red line indicates the common minimum number of individuals (similar effort), permitting richness comparison among habitats. Habitat acronyms: MGI = flooded gallery forest; MG = gallery forest; CL = “campo limpo”; CU = “campo úmido”; CS = “campo sujo”; CC = “campo cerrado” and CE = “cerrado s.s.”.



The hypothesis that gallery forest is the richest habitat of the Cerrado (e. g. Redford and Fonseca 1986; Vieira and Palma 2005) was not confirmed (Bezerra *et al.* 2009; Hannibal and Cáceres 2011; Carmignotto and Aires 2011; Santos-Filho *et al.* 2012). The results from this study showed similar richness in the gallery forest and along the open areas (arboreal dense savannas) which could be explained by the low percentage of forest environments in the Cerrado (little available area), and the absence of canopy traps,

leading to a mammalian assemblage poorly represented by forest species (see Rodrigues *et al.* 2002). However, studies using canopy traps in gallery forests failed to record additional species, on the other hand, species show stratification along the vertical axis, which may explain the high number of species present in the gallery forests (Hannibal and Cáceres 2010). Also, it is possible that the open areas, although less stratified vertically, can harbor equivalent richness via horizontal heterogeneity (see August 1983; Colli *et al.* 2002). The Cerrado diversity, however, is a reflection of the combined mosaic of habitats (see Carmignotto *et al.* 2012).

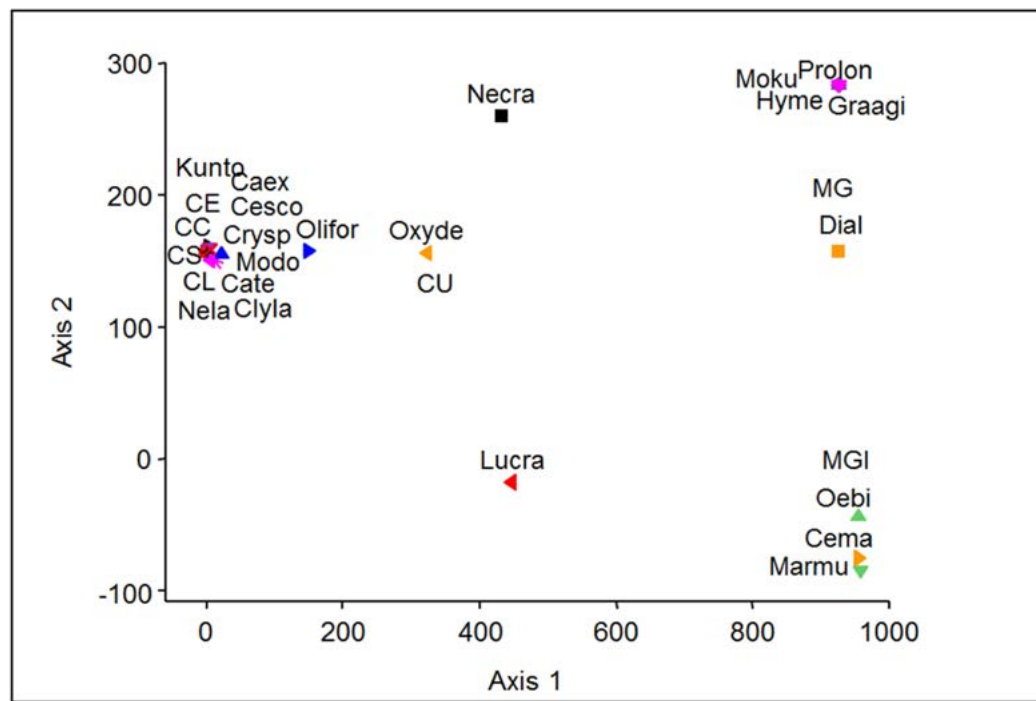


Figure 6. Scatterplot of the first and second axis DCA scores for the 20 species and the seven habitat types surveyed based in the capture success. The acronyms refer to the first genus and species syllabi and the habitats (CU = “campo úmido”; CL = “campo limpo”; CS = “campo sujo”; CC = “campo cerrado”; CE = “cerrado s. s.”; MGI= flooded gallery forest and MG = gallery forest).

Considering the species composition found in the region, there are some taxonomic comments that we also address. The specimens of *Cryptonanus* surveyed in ENP (cited as *Gracilinanus* sp. in Rodrigues *et al.* 2002) are morphologically distinct from the five currently valid species of *Cryptonanus* (Voss *et al.* 2005), representing an undescribed species for the genus. The only specimen of *Thylamys* recorded was *T. velutinus* (cited as *Thylamys* sp. in Rodrigues *et al.* 2002), which is an endemic Cerrado species with only few scattered records in the biome, and is considered vulnerable in the southern portion of the Cerrado (Carmignotto, 2009). Based on the karyotype data, ($2n = 52$ and $FNa = 52$ – diploid and autosomal fundamental numbers, respectively), the specimens of *Nectomys*, first considered as *N. squamipes* (Rodrigues *et al.* 2002), were correctly identified as *N. rattus* (see Bonvicino *et al.* 1996). Since Rocha *et al.* (2012) better delimited the taxonomic status of *Oecomys cleberi*, a member of the *O. bicolor* species group, and its geographic distribution extended to far southwestern Brazil, the specimens of *Oecomys* of the ENP might be referred to this taxon. However, there are great morphological overlap in this species complex, and due to absence of karyotype data for *O. cleberi* (ENP specimens showed $2n = 80$ and $FNa = 138$), we refer to these specimens as *O. bicolor* until molecular data are available for this population. Additionally, *C. marinhui* is morphologically very similar with *C. maracajuensis*, as are the karyotypes ($2n = 56$). We

classified the specimens surveyed at ENP as *C. marinhus* (cited as *Oryzomys gr. subflavus* sp. 2 in Rodrigues et al. 2002) based on the body and craniodental measurements that are closer to the mean for this taxon (see Percequillo et al. 2008), but molecular data also could aid in the taxonomic identity of these populations. The specimens referred to *Oryzomys gr. subflavus* sp. 1 in Rodrigues et al. (2002) showed similar morphology and karyotype (2n = 58 and FNa = 72) from that cited for *Cerradomys scotti* (see Percequillo et al. 2008).

Table 3. Distribution of the species encountered in the ENP showing the biomes of occurrence in Brazil and Central South America according to Paglia et al. (2012) and Carmignotto et al. (2012), respectively.

Species	Amazon	Atlantic Forest	Cerrado	Caatinga	Pantanal	Chaco	Pampas
<i>Chironectes minimus</i>	X	X	X		X		X
<i>Cryptonanus</i> sp.			X				
<i>Didelphis albiventris</i>			X	X	X	X	X
<i>Gracilinanus agilis</i>			X	X	X	X	
<i>Lutreolina crassicaudata</i>	X	X	X		X	X	X
<i>Marmosa murina</i>	X	X	X	X	X		
<i>Monodelphis domestica</i>			X	X	X	X	
<i>Monodelphis kunsii</i>			X		X	X	
<i>Philander opossum</i>	X		X	X	X		
<i>Thylamys velutinus</i>			X				
<i>Calomys expulsus</i>			X	X			
<i>Calomys tener</i>		X	X	X			
<i>Cerradomys marinhus</i>			X				
<i>Cerradomys scotti</i>			X		X	X	
<i>Hylaeamys megacephalus</i>	X	X	X		X		
<i>Kunsia tomentosus</i>			X				
<i>Necomys lasiurus</i>	X	X	X	X	X		
<i>Nectomys rattus</i>	X		X	X	X	X	
<i>Oecomys bicolor</i>	X		X		X		
<i>Oligoryzomys fornesi</i>			X	X	X	X	
<i>Oxymycterus delator</i>			X			X	
<i>Clyomys laticeps</i>			X		X	X	
<i>Proechimys longicaudatus</i>	X		X		X	X	
	9 (39%)	6 (26%)	23 (100%)	10 (43%)	16 (69%)	11 (48%)	3 (13%)

Other biomes adjoin the Cerrado and influence its faunal composition. The greater similarity of the Cerrado fauna with other biomes can be explained by the two components present in this biome: the forest and the open area physiognomies. Compared to the Amazon and Atlantic Rainforest fauna, the small mammals of the present study showed greater similarity with the Amazon, contrary to prior studies that found greater similarity with the Atlantic Rainforest (i.e. Redford and Fonseca 1986). This pattern is mainly related to the presence of four taxa that occur in both the Amazon and Cerrado, but are replaced by sister groups in the Atlantic Rainforest, such as *Philander frenatus* (not *P. opossum*), *Nectomys squamipes* (not *N. rattus*), *Oecomys catherinae* (not *O. bicolor*) and the genus *Trinomys* (not *Proechimys*). Because the systematics and taxonomic status of these groups are still being clarified previous studies did not realize these vicariant patterns between

the Amazon and Atlantic Rainforest fauna (see Costa 2003; Carmignotto *et al.* 2012).

Among the open areas, the Pantanal fauna is the most similar, followed by the Chaco (see Table 3). The presence of open and forested areas in these biomes provides potential habitats for the Cerrado small mammals. In addition, we found more species associated with the open areas than with the gallery forests. This pattern, shown also in other studies, contradicts the previous hypothesis that the fauna of the Cerrado and the Caatinga is, in general, associated with forest environments, with no specialization to xeric conditions and endemics restricted to the gallery forests (*i. e.* Redford and Fonseca 1986), demonstrating the high richness and specialization of mammal species to the open habitats (*e. g.* Hershkovitz 1993; Oliveira *et al.* 2003; Oliveira and Pessôa 2005).

Comparing the species composition of ENP with other areas surveyed in the Cerrado and transitional areas, we observed that there is a fauna more similar to the ones from the western portion of the Cerrado (Emmons 2009; Santos-Filho *et al.* 2012; Owen 2013), represented by species with a distribution restricted or more concentrated in this portion of the biome such as *Thylamys velutinus*, *Kunsia tomentosus*, *Clyomys laticeps* and *Proechimys longicaudatus*. The results obtained here reinforce the great variability among the nonvolant small mammal assemblages of the Cerrado and indicate that the survey of new areas can change the current thoughts about the mammalian fauna in this biome (see Carmignotto *et al.* 2012). The Cerrado is considered one of the 25 hotspots of world's biodiversity because the high level of endemic species and the high proportion of habitat destruction (Myers *et al.* 2000). Land conversion to crops and pastures has already occurred in most of the Cerrado, and there are few areas preserved and available for conservation efforts. The variability in species composition and in the structure of the small mammal communities also reflects the need to preserve different regions of the Cerrado biome to effectively maintain its biodiversity (Marinho-Filho *et al.* 1994; Vieira and Palma 2005).

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Specimens examined from the Emas National Park (ENP), Goiás state, Brazil. Voucher number acronyms: MN, Museu Nacional/Universidade Federal do Rio de Janeiro; MZUSP – Museu de Zoologia da Universidade São Paulo; UNB – Mammal collection of the Universidade de Brasília; and APC – the field number of Ana Paula Carmignotto, which will be deposited in the MN.

Order Didelphimorphia
Family Didelphidae Gray 1821

Cryptonanus Voss et al. 2005

Cryptonanus sp. ($n = 4$, skin, skull and skeleton): APC: 573, 593, and 595 (females); 582 (male).

Didelphis Linnaeus 1758

Didelphis albiventris Lund, 1840 ($n = 2$, skin and skull): MN: 71569 and UNB: 1704 (males).

Gracilinanus Gardner and Creighton 1989

Gracilinanus agilis (Burmeister 1854) ($n = 2$, skin, skull and skeleton): APC: 614 and 655 (females).

Lutreolina Thomas 1910

Lutreolina crassicaudata (Desmarest 1804) ($n = 4$, skin, skull and skeleton): MN: 71697 and UNB: 1927 (females); MN: 71673 and UNB: 1707 (males).

Marmosa Gray 1821

Marmosa murina (Linnaeus 1758) ($n = 9$, skin and skull): MN: 71648, 71658, 71669, UNB: 2548, 2551, and 2552 (females); MN: 71990, UNB: 2549, and 2550 (males).

Monodelphis Burnett 1830

Monodelphis domestica (Wagner 1842) ($n = 1$, skull and skin): UNB: 1708 (male).

Monodelphis kunsi Pine 1975 ($n = 1$, skull and skin): APC: 612 (male).

Thylamys Gray, 1843

Thylamys velutinus (Wagner 1842) ($n = 1$, skull and fluid): MZUSP: 32098 (male).

Order Rodentia
Family Cricetidae Fischer, 1817
Subfamily Sigmodontinae Wagner, 1843

Calomys Waterhouse 1837

Calomys expulsus (Lund 1841) ($n = 7$, skin and skull): MN: 71660, 71664, 71677, and UNB: 2090 (females); UNB: 2089 (male); MN: 71667 and UNB: 2091 (undetermined).

Calomys tener (Winge 1887) ($n = 4$, skin and skull): MN: 71661 and UNB: 2092 (females); MN: 71680 and UNB: 2093 (males).

Cerradomys Weksler et al. 2006

Cerradomys marinhos (Bonvicino 2003) ($n = 11$, skin and skull): UNB: 1912, 1916, 1918, MN: 71654, 71662, and 71689 (females); UNB: 1901, 1910, MN: 71649, 71653, and 71687 (males).

Cerradomys scotti (Langguth and Bonvicino 2002) ($n = 14$, skin and skull) = MN: 71646, 71671, 71698, UNB: 1904, and UNB: 1908 (females); MN: 71647, 71686, 71692, UNB: 1905, 1906, and 1915 (males); MN: 71666, 71709 (only skull), and UNB: 1922 (undetermined).

Hylaeamys Weksler et al. 2006

Hylaeamys megacephalus (Fischer 1814) ($n = 18$, skin and skull): MN 71683, 71684, 71704, 71705, UNB: 1909, 1907, 1913, and 1924 (females); UNB: 1921, MN: 71681, 71693, 71699, 71700, and 71701 (males); MN: 71685, UNB: 1914, 1919, 1920 (undetermined).

Kunsia Hershkovitz 1966

Kunsia tomentosus (Lichtenstein 1830) ($n = 4$, skin and skull): MN: 62567 (female); MN: 62569, UNB: 1705, 1706 (males).

Necomys Ameghino 1889

Necomys lasiurus (Lund 1841) ($n = 28$ skin and skull, and 10 fluid): UNB: 1690, 1691, 1692, 1700, 1701, 1702, MN: 68978, 68980, 68981, 68983, 68985, 68984, 71682, 71702, 71706, 71707, 71713 (females), UNB: 1693, 1694, 1695, 1696, 1698, 1699, 1703, MN: 68976, 68982, 68977, 71670, 71679, 71694, 71695, 71708, 71711, 71712, 71715, 71714 (males), UNB: 1928, MN: 68979 (undetermined).

Nectomys Peters 1861

Nectomys rattus (Pelzeln 1883) ($n = 5$, skin and skull): MN: 71663 and UNB: 1710 (females); MN: 71655, UNB: 1709, and 1711 (males).

Oecomys Thomas 1906

Oecomys bicolor (Tomes 1860) ($n = 17$, skin and skull): MN: 71691, 71703, UNB: 1923, 1925, and 1917 (females); MN: 71650, 71651, 71656, 71672,

71688, UNB: 1712, 1713, 1716, and 1903 (males); MN: 71665, and UNB: 1926 (undetermined).

Oligoryzomys Bangs 1900

Oligoryzomys fornesi (Massoia 1973) ($n = 2$, skin and skull): MN: 71654 and UNB: 2553 (males).

Oxymycterus Waterhouse 1837

Oxymycterus delator Thomas 1903 ($n = 14$, skin and skull): MN: 71674, 71675, 71697, 71710, UNB: 2082, 2085, 2087, 2088 (females); MN: 71652, 71657, 71678, UNB: 2083, 2084, 2086 (males).

Family Echimyidae Gray, 1825

Clyomys Thomas, 1916

Clyomys laticeps (Thomas 1909) ($n = 17$, 1 skull, 15 skin and skull, and 2 fluid): MN: 68162, 68167, 68169, UNB: 2150, 2152, 2154, 2155 (females); MN: 68163, 68164, 68165, 68166, 68168, 68172, 68173, UNB: 2151, and 2153 (males); MN: 68170 (undetermined).

Proechimys Allen, 1899

Proechimys longicaudatus (Rengger 1830) ($n = 3$, skin and skull): UNB: 1911 (female); MN: 71668, and UNB: 1902 (undetermined).